

The nutrient status of South African rivers: concentrations, trends and fluxes from the 1970s to 2005

S. de Villiers^{a,b*} and C. Thiert^{b,c}

Eutrophication of river systems, resulting from nutrient enrichment, is globally considered to be one of the most serious threats to freshwater ecosystem services such as water quality and biodiversity. This study provides a comprehensive overview of the nutrient status of the 20 largest river catchments in South Africa, based on dissolved inorganic nitrogen ($\text{NO}_3^- + \text{NO}_2^-$) and phosphorus (PO_4^{3-}) long-term water quality monitoring data collected by the Department of Water Affairs and Forestry. Nutrient levels exceeding recommended water quality guidelines for plant life are observed in all of the rivers, except one. Additionally, dissolved-phosphorus levels exceeding recommended concentrations for aquatic animal life prevail episodically in all but 6 of the catchments. Alarming, statistically significant ($P < 0.05$) upward trends in dissolved PO_4^{3-} levels are found in almost 60% of the rivers evaluated. The most likely cause of increasing nutrient enrichment is effluent from dysfunctional sewage works and unsewered human settlements. This poses a serious and costly threat to water quality and biodiversity. Nutrient fluxes associated with agricultural runoff, representing loss of soil fertility, translate into fertilizer-equivalent costs exceeding several hundred million rands annually.

Introduction

Anthropogenic disturbances to the natural nitrogen and phosphorus cycles over the last couple of decades have resulted in eutrophication being considered one of the most serious problems facing freshwater ecosystems globally.¹⁻⁶ Nutrient enrichment alters the competitive balance between plant species, resulting in the degradation of aquatic plant communities, which provides food, shelter and breeding habitats for a range of animal species. Additionally, elevated nutrient levels can be detrimental to the health of humans and toxic to aquatic animals. The most dramatic manifestation of eutrophication in freshwater and marine aquatic systems is extensive kills of both invertebrates and fishes, as a result of oxygen depletion related to the decomposition of the excess organic matter produced. Nutrient levels in rivers are also of significance to the health of water bodies fed by it, such as wetlands, estuaries, groundwater and surface freshwater reservoirs.

The dissolved inorganic nitrogen and phosphorus content of river water derives from both natural and anthropogenic sources. In the absence of pollution, the nutrient content of river water represents primarily a balance between the production of dissolved species through the chemical weathering of nitrogen- and phosphorus-containing geological deposits in the catchment, and consumption by biological productivity in the system.^{1,4,7}

The principal anthropogenic point sources of inorganic nitrogen and phosphorus in aquatic ecosystems are: municipal sewage effluents and overflows of storm and sanitary sewers, wastewater from livestock farming, industrial wastewater effluents, and runoff from waste disposal sites, working mines and unsewered industrial sites. The main anthropogenic diffuse sources are: agricultural activities (use of manure and nitrogenous fertilizers, cultivation of N_2 -fixing crops), runoff from nitrogen-saturated and burned forests and grasslands, urban runoff from unsewered, sewerage and failed septic systems, runoff from construction sites and abandoned mines, polluted ground waters, anthropogenic atmospheric deposition loads (such as from fossil fuel combustion) and biomass burning. Globally, the dominant source of the increase, by a factor of about 4, in nutrient levels are widespread agricultural intensification and increased discharge of domestic wastes.¹⁻⁴

This study provides the first comprehensive overview of the nutrient status of South Africa's rivers. In addition to geographic gradients in dissolved [$\text{NO}_3^- + \text{NO}_2^-$] and [PO_4^{3-}] levels, seasonal nutrient profiles, temporal trends and fluxes are evaluated. The classification of the trophic status of South Africa's aquatic ecosystems is currently restricted to four broad categories: oligotrophic, mesotrophic, eutrophic, and hypertrophic.^{8,9} Additionally, Target Water Quality Range (TWQR) values are stipulated for the water-use sectors of domestic, recreational, industrial, irrigation, stock watering, and aquaculture,⁹ but not for aquatic ecosystems. This is because aquatic ecosystems in different localities have very different requirements, and the site-specific studies needed to derive TWQR values have not been carried out in South Africa. River water nutrient levels presented in this study are, therefore, compared to internationally accepted guidelines for aquatic ecosystem health.

Study area, database and data handling

Long-term water quality monitoring results for the 20 largest primary river catchments in South Africa (see Table 1, and Table 2 and Fig. A in supplementary material online) was generously provided by the Department of Water Affairs and Forestry (DWAF) on special request. Water quality monitoring parameters include the dissolved inorganic nitrogen species [$\text{NO}_2^- + \text{NO}_3^-$] (expressed as $\mu\text{g N l}^{-1}$, with 20–40 $\mu\text{g N l}^{-1}$ detection limits) and dissolved inorganic phosphate (also referred to as soluble reactive phosphate, SRP) reported as [PO_4^{3-}] ($\mu\text{g P l}^{-1}$, with 3–5 $\mu\text{g P l}^{-1}$ detection limits) at all the stations evaluated. Data for [NH_4^+] and total dissolved phosphorus (TP) are available at only some of the stations or some sections of the record. Additionally, where [NH_4^+] data are available, values are typically close to or at the analytical detection limit ($\sim 40 \mu\text{g N l}^{-1}$) and less than 10% of [$\text{NO}_2^- + \text{NO}_3^-$] ($[\text{NO}_x]$) values. For the purposes of this study, therefore, the more comprehensive [NO_x] and [PO_4^{3-}] data only were evaluated.

Sampling frequency varies among sites, from almost weekly to monthly. Where more than one water quality data point is

^aDepartment of Geology, Geography and Environmental Studies, Stellenbosch University, Private Bag X1, Matieland 7602, South Africa.

^bAEON (African Earth Observatory Network).

^cDepartment of Statistical Sciences, University of Cape Town, Rondebosch 7700, South Africa.

*Author for correspondence. E-mail: steph@sun.ac.za

Table 1. Median, minimum and maximum values (of all monthly values in time series over the sampling period) for dissolved [$\text{NO}_3^- + \text{NO}_2^-$] and [PO_4^{3-}] and annual fluxes of N and P (calculated using the catchment areas given in Table 2 in supplementary material online.)

River catchment	Sampling period	[NO_3^-]($\mu\text{g N l}^{-1}$) Median[min; max]	[PO_4^{3-}]($\mu\text{g P l}^{-1}$) Median[min; max]	Annual flux	
				($\text{kg N km}^{-2} \text{ yr}^{-1}$)	($\text{kg P km}^{-2} \text{ yr}^{-1}$)
Lower Orange	1980–2000	104[40; 1180]	19[5; 106]	0.4	0.06
Upper Orange	1980–2004	542[40; 905]	22[5; 92]	11.1	0.47
Upper Vaal	1975–2005	127[20; 2110]	36[3; 272]		
Harts	1972–2004	209[40; 1802]	21[5; 88]	0.5	0.07
Riet	1987–2005	108[40; 550]	21[5; 441]	0.5	0.10
Wilge	1975–2005	220[40; 2140]	37[5; 213]	8.0	1.35
Olifants	1995–2002	102[20; 595]	13[3; 59]	24.8	2.49
Berg	1974–2005	378[40; 2434]	24[5; 487]	54.9	2.41
Breede	1973–2004	134[40; 1 450]	15[5; 96]	19.4	1.28
Gourits	1976–2005	35[20; 918]	32[3; 363]	0.1	0.10
Keurbooms	1977–2005	20[20; 350]	13[3; 80]	0.01	0.04
Gamtoos (Groot)	1972–2006	30[20; 1177]	17[3; 81]	0.02	0.01
Swartkops	1995–2005	418[20; 3085]	46[12; 626]	0.88	0.11
Sundays	1971–2004	117[40; 2030]	40[5; 335]	0.17	0.06
Great Fish	1971–2005	429[40; 2669]	78[5; 1172]	3.22	0.66
Keiskamma	1971–2005	342[20; 2630]	27[3; 403]	3.67	0.38
Great Kei	1990–2005	158[20; 1465]	26[3; 146]	2.97	0.51
Mzimvubu	1980–2005	179[20; 693]	24[3; 228]	17.7	3.38
Mkomazi	1978–2005	97[20; 1380]	15[3; 161]	23.6	3.05
Tugela	1977–2004	192[20; 4020]	25[5; 223]	15.9	1.61
Mfolozi	1995–2005	64[20; 3114]	26[5; 223]	3.91	1.56
Phongola	1972–2004	500[20; 3100]	18[5; 211]	46.2	2.20
Komati	1982–2005	370[20; 1343]	20[5; 192]	7.46	0.44
Olifants	1983–2005	293[20; 2800]	28[5; 189]	3.53	0.32
Limpopo	1993–2003	57[20; 1588]	27[5; 145]	0.24	0.13

available in any given month, an average value was calculated to provide time-series data at a monthly resolution. This also provides compatibility with the DWAF's total monthly water flow records (available at www.dwaf.gov.za). Further data reductions, such as the calculation of median concentration values for the entire time series and monthly medians for the construction of a representative annual profile for each site (Fig. 1) are elaborated on below.

The aim of this study was to assess comparative nutrient concentrations and fluxes at the primary catchment scale. To that effect, water quality monitoring stations as far downstream as possible in each catchment area were selected. An exception to this is the large Orange River system, which was also evaluated at a secondary catchment scale at monitoring sites along the Vaal River and its tributaries. A total of 25 coincident water quality and flow monitoring stations with drainage areas, ranging from 625 to more than 650 000 km^2 , were evaluated and statistically analysed (see Table 1 and Fig. A online). The cumulative area represented by the sum of the catchment areas of the chosen sampling points represents more than 95% of the surface area of South Africa.

The damming of rivers and the construction of water transfer pipelines has had a dramatic effect on water budgets within South Africa's river systems. The most significant of these are the completion of the Vaal Dam in 1938, the Bloemhof Dam (also on the Vaal River) in 1970, the Gariiep Dam in 1972, the Van der Kloof Dam in 1977 and the Orange–Fish Tunnel in 1975.¹⁰ The purpose of this study was not to evaluate the effect of these disturbances of water flow on nutrient budgets. In almost all instances, in fact, water quality monitoring data represent conditions after the dam building phase of the early 1970s. Notable exceptions are the Upper Vaal and Fish rivers. Extensive damming of the Upper Vaal prohibits combining water flow data with nutrient concentrations as a meaningful way of calculating fluxes in the present

Upper Vaal River catchment. The Orange–Fish Tunnel water transfer scheme has similar implications for nutrient levels and fluxes in the Fish and Sundays rivers, and data presented have to be interpreted in that context. In some instances, long-term water quality monitoring is carried out at sites at, or just downstream of, large dams, which does require the use of river flow data from upstream locations for meaningful flux calculation purposes. This is the case at the following water quality monitoring sites (water flow station in brackets): the Van der Kloof Dam (Vluytjieskraal at 29.809°S, 24.438°E), Harts River (Espagsdrift upstream of the Loskop Dam at 27.902°S, 24.609°E), and Riet River (Kromdraai upstream of the Kalkfontein Dam at 29.655°S, 25.970°E).

Long-term temporal trends in nutrient concentrations were statistically evaluated by fitting a straight line to the data, with the method of ordinary least squares (OLS). The OLS procedure provides a best linear unbiased estimator of the slope and intercept if the Gauss–Markov conditions are met.¹¹ Some of the data sets exhibit serial correlation or autocorrelation, in which case the slope was estimated using a generalized least-squares estimator according to the Prais–Winstone method.¹²

Results and discussion

River water nitrate, nitrite and phosphate levels

Median [$\text{NO}_3^- + \text{NO}_2^-$] and [PO_4^{3-}] values, together with the minimum and maximum values in the time series, are listed in Table 1. Nutrient levels close to or below the analytical detection limits are observed in all of the rivers at some stage during the observation period (minimum values in Table 1). These values, $<40 \mu\text{g N l}^{-1}$ and $<5 \mu\text{g P l}^{-1}$, are indicative of near pristine or low natural background levels. Recommended water quality criteria for the protection of aquatic animals are 80–350 $\mu\text{g NO}_2^- \text{ N l}^{-1}$ and 2000–3600 $\mu\text{g NO}_3^- \text{ N l}^{-1}$ for the inorganic nitrogen in the form of

nitrate and nitrite,^{13,14} and between 20 and 100 $\mu\text{g P l}^{-1}$ for soluble reactive phosphorus.⁶ Unionized ammonia (NH_3) is the most toxic form of inorganic nitrogen to aquatic animals, and water quality criteria ranging from 50–350 $\mu\text{g NH}_3\text{-N l}^{-1}$ for short-term exposures and 10–20 $\mu\text{g NH}_3\text{-N l}^{-1}$ for long-term exposures have been recommended.^{5,15–17} Evaluation of available pH and NH_4^+ data for South African rivers suggests that, at the primary catchment scale, NH_3 concentrations are negligible and not a concern (de Villiers, unpublished).

Recommended levels of dissolved inorganic nitrogen and phosphorus, for the prevention of eutrophication, are lower than those listed above for aquatic animals. There is some uncertainty about the lower limit of phosphorus required for freshwater plant growth. Levels higher than 30 $\mu\text{g total P l}^{-1}$ is generally considered conducive to eutrophication, provided that inorganic nitrogen or other nutrients are not limiting.^{5,18} Plants require nitrogen and phosphorus in a ratio of between 7 and 8 (weight ratio) and concomitant dissolved values of > 400 $\mu\text{g total N l}^{-1}$ and > 30 $\mu\text{g total P l}^{-1}$ are generally considered favourable for eutrophication in freshwater systems. Dissolved $[\text{NO}_x]$ accounts for most of the total dissolved nitrogen, but dissolved $[\text{PO}_4^{3-}]$ is typically only a fraction of total phosphorus. Available total phosphorus data for the Great Fish, Great Kei and Olifants (Mpumalanga) rivers suggest total phosphorus/ PO_4^{3-} loads of 1.9, 6.8 and 3.2 $\mu\text{g P l}^{-1}$, respectively (unpublished data), i.e. dissolved $[\text{PO}_4^{3-}]$ levels account for only a fraction of the total phosphorus present. For the purposes of the discussion below, $[\text{PO}_4^{3-}] = 20 \mu\text{g P l}^{-1}$ is conservatively used as a threshold value for eutrophication in freshwater systems, in combination with the recommended 400 $\mu\text{g N l}^{-1}$ value for $[\text{NO}_x]$. One should note, however, that aquatic animals naturally adapted to low inorganic nitrogen levels may have lower toxicity thresholds and that these values are unknown for South Africa's freshwater ecosystems.

Median $[\text{NO}_x]$ values exceeding 400 $\mu\text{g N l}^{-1}$ are found in the Swartkops, Phongola, Upper Orange and Great Fish rivers (Table 1). Seasonal $[\text{NO}_x]$ profiles (Fig. 1f, j, o, s) confirm the occurrence of values exceeding 400 $\mu\text{g N l}^{-1}$ throughout, or most of, the year in these catchments. However, maximum values for $[\text{NO}_x]$ in the time-series data for each catchment indicate that, with the exception of the Keurbooms River, values exceeding 400 $\mu\text{g N l}^{-1}$ occur at least episodically in all of the river systems. The constructed seasonal profiles (Fig. 1) indicate that, in addition to the four rivers already mentioned, $[\text{NO}_x] > 400 \mu\text{g N l}^{-1}$ prevails for at least five months of the year in the Berg (Fig. 1a) and Komati (Fig. 1k) rivers and for one to two months a year in the Upper Vaal (Fig. 1e), Olifants (Mpumalanga) (Fig. 1h), Limpopo (Fig. 1m) and Keiskamma (Fig. 1u) river systems. The Gourits (Fig. 1v), Gamtoos (Fig. 1w) and Keurbooms (Fig. 1x) rivers are the most pristine in respect of their $[\text{NO}_x]$ levels.

Median $[\text{PO}_4^{3-}]$ values exceeding 20 $\mu\text{g P l}^{-1}$ are found in 18 of the 25 catchments studied (Table 1). Additionally, maximum values in the time-series data suggest that concentrations exceed the 100 $\mu\text{g P l}^{-1}$ recommended maximum value for aquatic animal life⁶ at least episodically in all but six of the catchments (Table 1). Seasonal profiles reveal an even starker picture (Fig. 1; $10 \times [\text{PO}_4^{3-}]$ values plotted on the left-hand y-axis). $[\text{PO}_4^{3-}]$ values exceeding 20 $\mu\text{g P l}^{-1}$ prevail throughout the year in the Upper Vaal (Fig. 1e), Swartkops (Fig. 1f), Great Fish (Fig. 1s) and Gourits (Fig. 1v) river catchments, and for most of the year in the Berg (Fig. 1a), Tugela (Fig. 1c), Wilge (Fig. 1d), Olifants (Mpumalanga) (Fig. 1h), Phongola (Fig. 1j), Komati (Fig. 1k), Limpopo (Fig. 1m), Harts (Fig. 1n), Upper Orange (Fig. 1o), Mfolozi (Fig. 1p), Riet (Fig. 1q), Mzimvubu (Fig. 1r), Great Kei (Fig. 1t) and Keiskamma

(Fig. 1u) river catchments.

As mentioned above, the development of eutrophic conditions requires also the presence of nutrients and, provided nothing else is limiting plant productivity, $[\text{NO}_x] > 400 \mu\text{g N l}^{-1}$ and $[\text{PO}_4^{3-}] > 20 \mu\text{g P l}^{-1}$. Such conditions are observed for most of the seasonal cycle in the Berg (Fig. 1a), Swartkops (Fig. 1f), Komati (Fig. 1k) and Great Fish (Fig. 1s) rivers. The following catchments exhibit potential for eutrophication during part of the annual cycle: Upper Vaal (Fig. 1e), Olifants (Mpumalanga) (Fig. 1h) and Keiskamma (Fig. 1u) rivers. It is important to note that evaluation of data at the primary catchment scale provides only a conservative estimate of the potential threat of eutrophication. At smaller geographic scales, for example in the close vicinity of point sources, much greater levels of nutrient enrichment will occur.

Temporal trends

Very significant ($P < 0.05$) downward trends in $[\text{NO}_x]$ are observed in 7 of the catchments studied (see Table 2 in supplementary material online): the Upper Orange, Upper Vaal, Harts, Sundays, Keiskamma, Tugela and Olifants (Mpumalanga) rivers. With the exception of the Keiskamma system, all of these catchments support substantial agricultural activity and the downward trend in $[\text{NO}_x]$, and elevated $[\text{NO}_x]$ are consistent with reduced use of fertilizers since the late 1970s.¹⁹ The intensively cultivated Berg and Breede river systems, however, show signs of increasing $[\text{NO}_x]$ levels, at the $P < 0.10$ level in the case of the Breede (see Table 3 online). This is consistent with enhanced agricultural activity in their particular catchments, in contrast to the national trend. There is also evidence for increasing $[\text{NO}_x]$ levels, at $P < 0.10$, in the Riet and Phongola systems.

A significant downward trend in dissolved $[\text{PO}_4^{3-}]$ ($P < 0.001$) is found in only one catchment, the Upper Orange River (Table 3), associated with a significant downward trend in $[\text{NO}_x]$. Fourteen of the other 24 monitoring sites manifest significant upward trends in $[\text{PO}_4^{3-}]$, at $P < 0.05$, and an additional two sites an upward trend at $P < 0.10$ (Table 3). The most pronounced increase in $[\text{PO}_4^{3-}]$ is found in the Swartkops River, with levels increasing at 10.05 $\mu\text{g P l}^{-1} \text{ yr}^{-1}$. With a median value of 46 $\mu\text{g P l}^{-1}$, this translates into a $[\text{PO}_4^{3-}]$ doubling time of less than 5 years. In combination with the high median and rising $[\text{NO}_x]$ levels, the Swartkops presents itself as one of the most threatened freshwater systems in South Africa.

Discrete versus point sources of nutrients

Diffuse nutrient sources produce seasonal concentration profiles coincident with river runoff, that is, concentrations that peak during high runoff conditions. Classic examples of the concentration–runoff profiles produced by a dominant diffuse source are observed for $[\text{NO}_x]$ and $[\text{PO}_4^{3-}]$ in the Berg (Fig. 1a), Breede (Fig. 1b), Tugela (Fig. 1c), Wilge (Fig. 1d), Upper Vaal (Fig. 1e) and Swartkops (Fig. 1f) rivers. Agricultural activity is an important use of land (Table 1) in all of these areas and the seasonal nutrient profiles are therefore consistent with fertilizer application as the primary diffuse source of both $[\text{NO}_x]$ and $[\text{PO}_4^{3-}]$ in these catchments. A dominant diffuse source is also suggested by seasonal $[\text{NO}_x]$ profiles in the Olifants (W. Cape) (Fig. 1g), Olifants (Mpumalanga) (Fig. 1h) and Mkomazi (Fig. 1i) rivers, and seasonal $[\text{PO}_4^{3-}]$ profiles in the Olifants (Mpumalanga) (Fig. 1h), Limpopo (Fig. 1m), Harts (Fig. 1n) and Mzimvubu (Fig. 1r) rivers.

Point sources, in contrast to diffuse sources, result in seasonal concentration profiles that have no relation to runoff they provide a relatively constant input throughout the year, or have an inverse relation to river runoff. Catchments that demonstrate

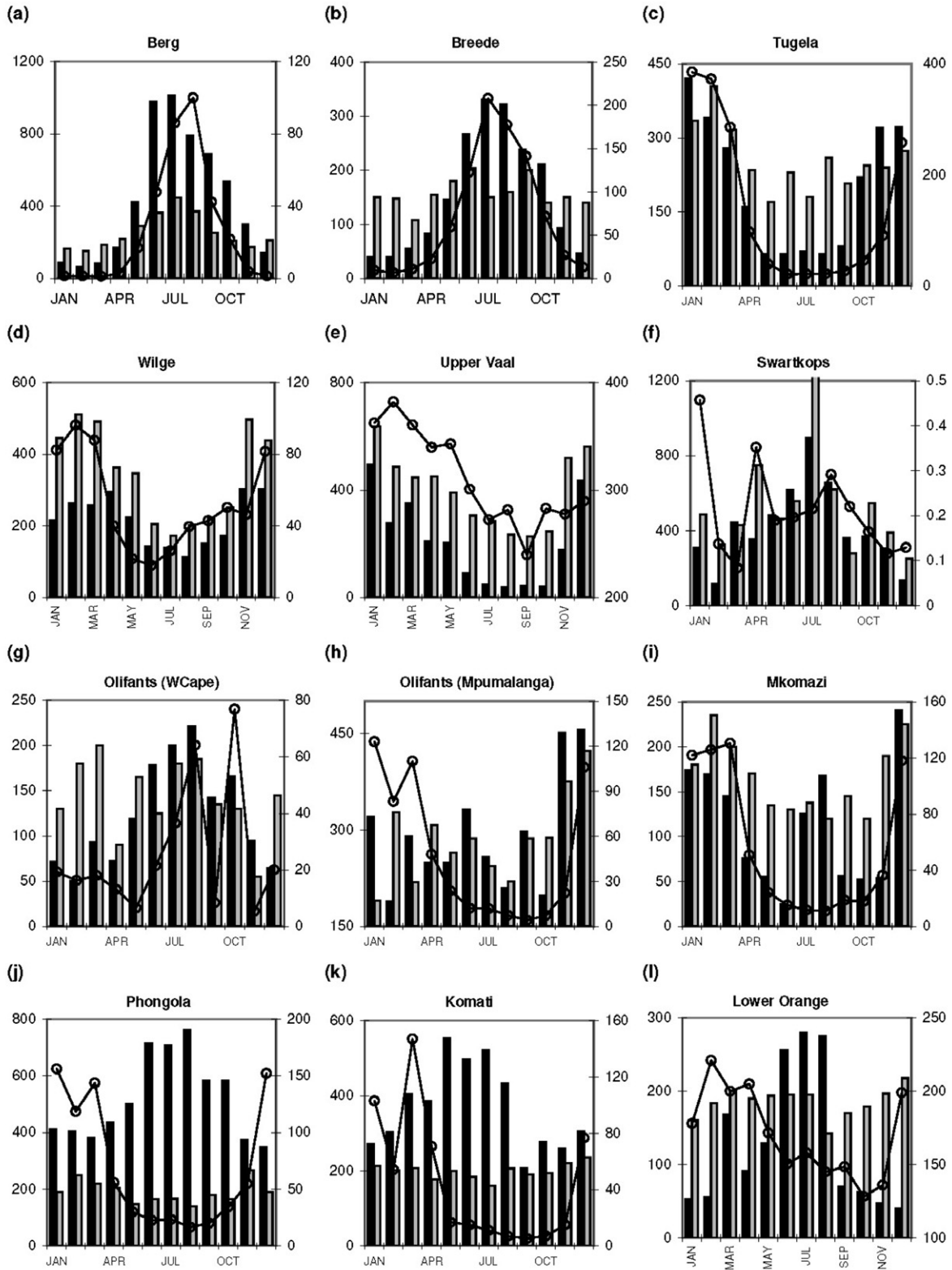


Fig. 1. Seasonal profiles (from monthly median values) of [NO_x] (black bars, left-hand y-axis in μg N l⁻¹, [PO₄³⁻] (grey bars, left-hand y-axis in 10 × μg P l⁻¹), and river flow (line, right-hand y-axis in 10⁶ m³) for river monitoring stations. *Figure continued page 347.*

clear evidence for a dominant point source for elevated dissolved [NO_x] (>400 μg N l⁻¹) are: the Phongola (Fig. 1j), Komati (Fig. 1k), Lower Orange (Fig. 1l), Limpopo (Fig. 1m), Harts (Fig. 1n) and Upper Orange (Fig. 1o) rivers. Strong evidence for a dominant point source of elevated [PO₄³⁻] (>20 μg P l⁻¹) is found in the Gourits (Fig. 1v) River only. Sewage pollution, from either dysfunctional wastewater treatment

plants or unsewered human settlements, is the main point source of both [NO_x] and [PO₄³⁻] in river systems.

The Phongola and Komati rivers have seasonal [NO_x] profiles that, as mentioned above, indicate a dominant point source. 'Background' concentrations during high runoff conditions, however, are quite high in both these rivers. It is possible, considering the existence of sugarcane plantations and dry-season